



# Karstic Microrefugia Host Functionally Specific Ant Assemblages

Zoltán Bátori<sup>1</sup>\*, Gábor Lőrinczi<sup>1</sup>, Csaba Tölgyesi<sup>1</sup>, Gábor Módra<sup>1,2</sup>, Orsolya Juhász<sup>1,3</sup>, Dianne Joy Aguilon<sup>1,2,4</sup>, András Vojtkó<sup>5</sup>, Orsolya Valkó<sup>6</sup>, Balázs Deák<sup>6</sup>, László Erdős<sup>7</sup> and István Elek Maák<sup>1,8</sup>

<sup>1</sup> Department of Ecology, Institute of Biology, University of Szeged, Szeged, Hungary, <sup>2</sup> Doctoral School of Environmental Sciences, University of Szeged, Szeged, Hungary, <sup>3</sup> Doctoral School in Biology, University of Szeged, Szeged, Hungary, <sup>4</sup> Department of Forest Biological Sciences, College of Forestry and Natural Resources, University of the Philippines Los Baños, Laguna, Philippines, <sup>5</sup> Department of Botany, Eszterházy Károly University of Applied Sciences, Eger, Hungary, <sup>6</sup> Lendület Seed Ecology Research Group, Institute of Ecology and Botany, Centre for Ecological Research, Vácrátót, Hungary, <sup>7</sup> Institute of Ecology and Botany, Centre for Ecological Research, Vácrátót, Hungary, <sup>8</sup> Museum and Institute of Zoology, Polish Academy of Sciences, Warsaw, Poland

#### **OPEN ACCESS**

#### Edited by:

Fernando Augusto Oliveira Silveira, Federal University of Minas Gerais, Brazil

#### Reviewed by:

Fabricio Beggiato Baccaro, Federal University of Amazonas, Brazil Francisco Emmanuel Mendez Castro, Institute of Botany (ASCR), Czechia

> \*Correspondence: Zoltán Bátori zbatory@gmail.com

#### Specialty section:

This article was submitted to Biogeography and Macroecology, a section of the journal Frontiers in Ecology and Evolution

Received: 03 October 2020 Accepted: 07 December 2020 Published: 23 December 2020

#### Citation:

Bátori Z, Lőrinczi G, Tölgyesi C, Módra G, Juhász O, Aguilon DJ, Vojtkó A, Valkó O, Deák B, Erdős L and Maák IE (2020) Karstic Microrefugia Host Functionally Specific Ant Assemblages. Front. Ecol. Evol. 8:613738. doi: 10.3389/fevo.2020.613738 Karst landscapes are among the topographically most complex systems with various microhabitats, where species can persist despite unfavourable macro-environmental changes. These microhabitats can also function as stepping stones during range shifts. Although the enclosed depressions (dolines, sinkholes or tiankengs) of karst landscapes may act as such safe havens, data on the functional diversity of their animal assemblages are scarce. Here, we investigate the functional diversity (i.e., certain functional groups and functional traits) of ant assemblages in dolines and study whether dolines surrounded by resource-poor environments (i.e., Fagus sylvatica forests) may function as safe havens for different kinds of ants. We found that dolines have the potential to maintain distinctive ant assemblages characterised by specific functional groups and traits that are rare in the surrounding habitats. Although continuous Fagus sylvatica cover in dolines had a detrimental impact on ant assemblages, grassland dolines surrounded by grasslands or Fagus sylvatica forests supported the presence of some specific functional groups and traits. These results suggest that conservation management needs to consider the influence of vegetation characteristics not only in dolines but also on the surrounding plateau. Moderate grazing and/or mowing would be desirable in order to prevent shrub encroachment into grasslands to ensure optimal vegetation structure for ants in the long run. Therefore, proper management and conservation of these safe havens may mitigate the rate of biodiversity loss under global warming. There is a need to explore a wide variety of taxonomic groups and taxonspecific traits in parallel with the quality of the surrounding habitats when evaluating current and potential microrefugia.

Keywords: doline, functional group, functional trait, hostile environment, microhabitat, refugia, safe haven, tiankeng

# INTRODUCTION

Microrefugia are small areas that retain locally favourable environments where species can survive during regional environmental changes (Keppel et al., 2012, 2015; Gentili et al., 2015). Although microrefugia are generally taxonspecific (Stewart et al., 2010), those that maintain the populations of various taxa are particularly important from a biogeographic and conservation point of view. There has recently been substantial progress in determining the location of such areas, demonstrating the importance of environmental heterogeneity (Ashcroft, 2010; Dobrowski, 2010; de Aguiar-Campos et al., 2020) and interactions between geomorphology, climatic change and the ecological tolerances of species (Bátori et al., 2017; Gentili et al., 2020; Michalak et al., 2020). Recent studies also emphasise the need to integrate functional trait analyses into the study of refugia, highlighting that more data on functional diversity from these areas are needed (Keppel et al., 2018; Ottaviani et al., 2020).

The potential of different landscapes to maintain high habitat diversity largely depends on the bedrock type and its physical and chemical properties. For instance, dissolution processes on limestone surfaces contribute to the formation of karst landscapes, covering nearly 20% of the Earth's land surface, and constituting one of the topographically most complex terrains (White et al., 1995; Culver, 2000; Bátori et al., 2020). These landscapes support several habitats, such as caves, limestone pavements, valleys, and enclosed depressions (dolines, sinkholes or tiankengs), where species composition and diversity vary with environmental heterogeneity (Whiteman et al., 2004; Bátori et al., 2009; Mammola et al., 2019). Botanists have long recognised the valuable role of these habitats as ideal natural laboratories in which to study multiple environmental gradients, thermal stability and their effects on plant species composition and vegetation patterns. Studies have shown that dolines may provide microhabitats for unique plants that are rare or absent from the surrounding landscape (Beck von Mannagetta, 1906; Horvat, 1953; Kobal et al., 2015), further reinforcing the idea that these depressions can function as habitat islands (Özkan et al., 2010; Bátori et al., 2014, 2017; Su et al., 2017). In addition, the accumulation of cold-adapted plants (e.g., high montane and glacial relict species) within certain microhabitats of dolines - such as poleward-facing slopes and bottoms - is a prime indicator of the presence of current warm-stage microrefugia. Although recent investigations have indicated that dolines may also provide safe havens for various functional groups of animals from different phyla (Vilisics et al., 2011; Kemencei et al., 2014; Raschmanová et al., 2015, 2018; Růžička et al., 2016; Battisti et al., 2017; Bátori et al., 2019), data on many taxa is scarce or completely lacking. Furthermore, most of these studies do not provide information about the effects of habitat heterogeneity and related vegetation patterns on the functional diversity of animal assemblages within dolines. In this paper, we focus on the distribution patterns of ants in habitats inside and outside of

forested and grassland dolines in a mountainous landscape in northern Hungary.

Ants are particularly sensitive to changes in resources, moisture and temperature (Sanders, 2002; Dolek et al., 2009), which makes them reliable indicators of environmental changes (Andersen, 1997; Arnan et al., 2014; Gallé, 2017), ecosystem health and functioning (Folgarait, 1998; Ottonetti et al., 2006; Eldridge et al., 2020). Generally, the species richness of ants is positively associated with habitat diversity (cf. Andersen, 1986; Báldi, 2008; Hortal et al., 2009; Pacheco and Vasconcelos, 2012), which may also affect the functional diversity of their assemblages (Gallé, 2017; Heinze, 2017). Topographically complex areas may contain a higher diversity of microhabitats providing suitable nesting sites and a higher amount of exploitable resources for different functional groups of ants (Ribas et al., 2003; Fayle et al., 2013). Ant assemblages are also strongly affected by changes in vegetation cover and structure. For instance, the presence of grassland patches in forested landscapes may support the maintenance of high ant diversity and facilitates the persistence of several species that cannot survive in shady habitats (Dolek et al., 2009). In contrast, homogenous habitats (e.g., Fagus sylvatica forests with low resource availability) create unfavourable conditions for many ants (Dolek et al., 2009; Wiezik et al., 2015). Results of the abovementioned studies suggest that habitat heterogeneity and related vegetation patterns may influence the composition of ant assemblages in current and potential microrefugia by filtering their functional traits, resulting in a relationship between the functional diversity of assemblages and small-scale environmental conditions (cf. Guilherme et al., 2019).

Here, we investigate how certain functional groups and traits of ant assemblages change with habitat diversity (dolines vs. plateau) and forest cover (Fagus sylvatica cover). We hypothesized that (1) habitat heterogeneity within dolines has a positive effect on the functional diversity of ant assemblages by supporting some specific functional groups and traits that are rare on the plateau - and would otherwise be eliminated in a warming climate, and that (2) the increasing cover of resource-poor environments (e.g., Fagus sylvatica forests) may compromise the capacity of dolines to act as functional refugia for ants within karst landscapes. We predicted that the extreme microclimatic conditions (i.e., warm and dry microhabitats on south-facing slopes vs. cool and moist microhabitats on north-facing slopes) within dolines (Bátori et al., 2019) favour species adapted to very different environmental conditions, while less extreme conditions on the plateau favour species adapted to intermediate conditions. The island-like attributes (e.g., funnel-shaped geometry of dolines) and the presence of Fagus sylvatica forests on the plateau may reinforce the spatial isolation of grassland microhabitats within dolines, resulting in the coexistence of less aggressive species or species with different social structure (cf. Tavella et al., 2018). We believe that this study provides further evidence to the hypothesis that heterogeneous karst landscapes may

maintain functionally diverse animal assemblages and will play an important role in buffering the negative effects of global warming.

# MATERIALS AND METHODS

### **Study Area**

Our study area is located on the karst plateau of the Bükk Mountains in northern Hungary  $(48^{\circ}04'31'' \text{ N}, 20^{\circ}29'57'' \text{ E};$  altitude: between 780 and 950 m), within the Bükk National

Park (Figures 1a,b). Mean annual temperature is 6.3°C and mean annual precipitation is 800 mm (Dövényi, 2010). Solution dolines are characteristic landforms of the area, having high topographic complexity (Figures 1c,d) and unique microclimate (Bárány-Kevei, 1999). The north-facing slopes of dolines receive less insolation than bottoms, south-facing slopes and the plateau, and therefore are consistently cooler and moister than their surroundings. However, higher insolation and temperature can be observed on south-facing slopes compared with other slopes and the plateau, providing warm "habitat islands" for many species (Bátori et al., 2019). At night, cold-air pooling is a



FIGURE 1 | Study region, study sites and sampling design. (a) Location of Hungary within Europe. (b) Location of the study region in Hungary. (c) Location of the study sites (site 1–6) in and around dolines. Forest cover is indicated by grey shading (doline types 2–4). (d) Grassland doline surrounded by grasslands. (e) Fagus sylvatica forest on the plateau. (f) Set-up of bait stations in a cross-shaped pattern. (g) Ants feeding on a bait.

typical phenomenon in doline bottoms, creating cool and moist conditions (Lehner et al., 2017).

European beech (*Fagus sylvatica*) is the dominant tree species within the study area (**Figure 1e**), but dense Norway spruce (*Picea abies*) plantations also occur close to the investigated dolines. Non-forested dolines maintain diverse grassland communities, including dry rocky swards on the south-facing slopes, and mesic–wet meadows on the north-facing slopes and bottoms. Mesic and semi-dry meadows are the most common grassland types on the plateau (Vojtkó, 2001). The concentration of high montane and glacial relict plants (e.g., *Aconitum moldavicum*, *Bupleurum longifolium*, and *Dracocephalum ruyschiana*) on the bottom and north-facing slopes of these dolines indicates the presence of current warm-stage microrefugia (Bátori et al., 2017).

#### **Sampling Design**

According to the vegetation cover (grassland vs. forest) inside and surrounding (i.e., plateau) the dolines, four doline types were identified as follows: type 1 - grassland dolines surrounded only by grasslands (forest cover in their surroundings: 0%), type 2 - grassland dolines surrounded by low amounts of forest (30-40%), type 3 - grassland dolines surrounded by large amounts of forest (90-100%), and type 4 - forested dolines surrounded only by forests (100%) (Figure 1c). Forested dolines surrounded by different amounts of grassland were absent from the study area. Three large solution dolines (diameter: 80-100 m, depth: 10-15 m) were selected from each type (12 dolines in total) to assess the species composition and relative abundance of ants. Six sampling sites were selected for each doline (72 sites in total), one on the north-facing slope, one in the bottom, one on the southfacing slope, and three on the surrounding plateau. The sites were located at least 20 m from each other. In types 2 and 3, the number of forest and grassland sites on the plateau corresponded to the relative proportions of the two habitat types.

The species composition and relative abundance of ants were assessed by non-destructive sampling methods such as baiting and hand collecting (cf. Bátori et al., 2019). Five bait stations were established at each site in a cross-shaped pattern at 5-m intervals (360 bait stations in total: 220 bait stations in grasslands, and 140 bait stations in forests) (Figures 1f,g and Supplementary Table 1). We used plastic discs (8 cm in diameter) with a quarter-teaspoon of tuna and honey mixture (1:3) as baits, and observed the foraging activity of ants every 40 min in five observation periods; the first observation started at 7:00 a.m., while the last started at 9:40 a.m. (overlapping with the daily peak activity period of ants). The presence and number of individuals belonging to different ant species were recorded at the baits during each observation period. Additionally, hand collecting was also employed to validate the species identity of occurring ants and to register those species that potentially avoided the baits. During hand collecting, we visually searched the ground surface at each site for 5 min, collecting any individuals (workers, gynes, etc.) found. Sampling was carried out in August 2018 in clear weather.

Ants were identified to morphospecies or species level whenever possible in the field, and representatives were collected

and preserved in 95% ethanol for later species validation. All the collected specimens were identified using the keys of Czechowski et al. (2012) and Seifert (2018), and deposited at the Department of Ecology, University of Szeged. Nomenclature follows the Bolton's catalog (Bolton, 2020) and the Hymenoptera Name Server (Johnson, 2007).

### **Species Grouping and Data Analysis**

Before analysis, all sampled ant species (22 species in total) were classified into four main functional groups that relate to (1) temperature preference, (2) moisture preference, (3) habitat preference, and (4) habitat plasticity (Czechowski et al., 2012; Seifert, 2018). Species were also classified according to three main functional traits that relate to: (1) dispersal ability, (2) aggressiveness, and (3) social structure (i.e., monogynous or polygynous colonies) (**Table 1** and **Supplementary Table 2**). Because the number of species and species occurrences were low in all forested sites (only one species, *Myrmica ruginodis*, was found at only eight baits), type 4 dolines were excluded from the statistical analyses.

The diagnostic species of habitats (doline vs. plateau) and doline types (types 1–3) were determined by fidelity calculations using the phi ( $\Phi$ ) coefficient of association, which ranges from –1 to + 1 (Tichý and Chytrý, 2006). Calculations were based on data from baits, and species with  $\Phi > 0.1$  were considered diagnostic (Fisher's exact test, p < 0.05). The highest value of phi was 0.3 in our case (see section "Results"). The value of + 1 would mean that the ant species occurred on all baits of the target habitat or doline type and was absent from the baits of the other habitat or

TABLE 1 | Functional groups and traits used in the study.

Functional groups and traits	Levels
	1: species adapted to warmer conditions
Temperature preference	2: species adapted to intermediate temperature conditions
	3: species adapted to cooler conditions
	1: species adapted to drier conditions
Moisture preference	2: species adapted to intermediate moisture conditions
	3: species adapted to moister conditions
	1: species associated primarily with grassland habitats
Habitat preference	<ol> <li>habitat generalists (species associated with forest and grassland habitats)</li> </ol>
	3: species associated primarily with forest habitats
	1: polytopic species
Habitat plasticity	2: oligotopic species
	3: stenotopic species
	1: aggressive species
Aggressiveness	2: moderately aggressive species
	3: subordinate species
	1: mainly polygynous species
Social structure	2: both polygynous and monogynous species
	3: mainly monogynous species
Dispersal ability	1: flying ability: good
	2: flying ability: poor

doline types. Fidelity measures were calculated using the JUICE program (Tichý, 2002).

Because many baits were visited by only one or a few species, we focused on the distribution of single functional groups and traits and did not use multi-trait approaches. We used generalized linear mixed-effects models (GLMMs) to analyse the effects of habitat (doline vs. plateau) and doline type (types 1–3) on the functional groups and traits of ants visiting the baits. The number



of species occurrences belonging to each category of the studied functional groups and traits was the response variable, while site was used as a random factor. We used a Poisson error distribution (maximum likelihood fit), but in some cases, where 0 and 1 scores predominated the dataset, we transformed the data to binary scale and used a binomial error distribution. GLMMs were performed in R (R Development Core Team., 2017) using the *glmer* function of the "lme4" package (Bates et al., 2013). Full models were tested for significance with the *Anova* function of the "car" package (Fox and Weisberg, 2011). We used the false discovery rate (FDR) method (*P. adjust* function) to account for multiple comparisons when performing pairwise comparisons.

### RESULTS

### **Diagnostic Species**

A total of 22 ant species were found in the study area (**Supplementary Table 2**). Baiting yielded 21 species, while 15 species were obtained by hand collecting. *Lasius platythorax* (fidelity value: 0.18), *Myrmica ruginodis* (0.23), *M. sabuleti* (0.27), and *Tetramorium* cf. *caespitum* (0.16) were diagnostic for dolines, while *L. niger* (0.24) and *M. scabrinodis* (0.30) for the plateau. All doline types (types 1–3) had their diagnostic species: *Tapinoma erraticum* (fidelity value: 0.24), *Formica pratensis* (0.21), and *Tetramorium* cf. *caespitum* (0.26) for types 1, 2 and 3, respectively.

# **Functional Group and Trait Composition**

We did not find significant difference between the total number of species in dolines and the plateau (**Table 2**). However, dolines and the plateau differed significantly from each other with respect to many functional groups and traits (**Figure 2**). Dolines had a significantly higher number of species adapted to cooler, warmer, drier or moister conditions, while the plateau had more species adapted to intermediate temperature and intermediate moisture conditions. The number of species associated primarily with forest habitats was higher in dolines, while the number of species associated primarily with grassland habitats was higher on the plateau. Dolines had more moderately aggressive and both polygynous and monogynous species, while the number of subordinate species was higher on the plateau (**Table 2**).

The total number of species and most functional groups and traits did not differ significantly among the different doline types (**Figure 3**, **Table 3**, and **Supplementary Table 3**). However, the number of species adapted to intermediate moisture conditions was higher in type 3 than in type 1, and types 2 and 3 contained more aggressive and monogynous species than type 1.

Hand collecting confirmed the results of baiting. For instance, the proportion of species adapted to cooler (34.5%), warmer (52.7%), drier (52.7%) or moister (7.3%) conditions was much higher in dolines, while the plateau contained a higher proportion of species adapted to intermediate temperature (42.3%) or moisture (65.4%) conditions. Generally, only small differences in functional group and trait proportions were observed among the different doline types.

 TABLE 2 | Comparisons of species numbers and ant species occurrences related

 to different functional groups and traits (see Figure 2 and Table 1) in different

 habitats (doline vs. plateau) in the Bükk Mts (Hungary) using the fitted generalized

 linear mixed-effect models.

Species number, functional groups and traits	Model	
	χ <sup>2</sup>	p
Total number of species	0.77	0.381
Warmer	6.95	0.008
Intermediate temperature	21.29	<0.001
Cooler	5.17	0.023
Drier	7.29	0.007
Intermediate moisture	17.62	<0.001
Moister	10.09	0.001
Grassland	4.13	0.042
Habitat generalist	0.02	0.882
Forest	14.81	<0.001
Polytopic	0.12	0.732
Oligotopic	4.67	0.031
Stenotopic	0.04	0.841
Aggressive	0.01	0.948
Moderately aggressive	30.75	<0.001
Subordinate	6.49	0.011
Polygynous	0.72	0.396
Polygynous, Monogynous	14.71	<0.001
Monogynous	1.87	0.178
Flying: good	0.41	0.523
Flying: poor	0.49	0.485

Significant differences are indicated by bold p values.

# DISCUSSION

Dolines provide microhabitats for a number of ant functional groups and traits that are rare in the surrounding habitats. This highlights the potential of dolines to function as safe havens during various environmental changes. We further demonstrated that continuous *Fagus sylvatica* cover in dolines had a detrimental impact on the species composition of ant assemblages, drastically reducing the number of functionally different species. However, higher *Fagus sylvatica* cover on the plateau did not hinder the functional diversity in grassland dolines.

Our findings underline the importance of dolines as local biodiversity hotspots in karst landscapes by supporting specific functional group patterns of ant assemblages. Grassland dolines in the study area acted as key habitats for ants adapted to cooler and/or moister conditions (e.g., *Myrmica lobicornis, M. ruginodis*, and *Lasius platythorax*), while species on the plateau indicated intermediate temperature and/or moisture conditions (e.g., *L. bombycina, L. niger*, and *M. scabrinodis*). This is due to the high habitat heterogeneity within dolines, introducing great variation in microclimates and soil moisture (Whiteman et al., 2004; Raschmanová et al., 2013; Bátori et al., 2019). Previous microclimatic studies indicated north-facing slopes of grassland dolines to be cooler and more humid than south-facing slopes, bottoms and plateaus during daytime, while cool-air pooling may provide higher humidity and extremely low temperatures



Ants in Karst Depressions

**TABLE 3** Comparisons of ant species occurrences related to different functional groups and traits (see **Figure 3** and **Table 1**) in doline types (types 1–3) in the Bükk Mts (Hungary) using the fitted generalized linear mixed-effect models (only statistically significant models are shown).

Model	Functional groups and traits Intermediate moisture		
Full model	χ <sup>2</sup>	p	
	6.30	0.043	
Pairwise comparisons	z	p	
type 1 vs. type 2	1.70	0.134	
type 1 vs. type 3	2.51	0.036	
type 2 vs. type 3	0.92	0.356	
	Aggressive		
Full model	χ <sup>2</sup>	p	
	21.47	<0.001	
Pairwise comparisons	z	p	
type 1 vs. type 2	2.56	0.027	
type 1 vs. type 3	3.91	<0.001	
type 2 vs. type 3	0.89	0.632	
	Monogynous		
Full model	χ <sup>2</sup>	p	
	8.44	0.015	
Pairwise comparisons	Z	p	
type 1 vs. type 2	2.56	0.028	
type 1 vs. type 3	2.76	0.016	
type 2 vs. type 3	0.24	0.968	

The p values were corrected with the false discovery rate (FDR) method. Significant differences are indicated by bold p values.

in doline bottoms at night (Bárány-Kevei, 1999). Although only a few studies have been published on the distribution of animal taxa in European dolines (and most of these studies focus on collapse dolines with cave entrances), their ability to support species adapted to cooler and/or moister conditions is widely accepted (Kemencei et al., 2014; Bátori et al., 2017; Su et al., 2017). For instance, Růžička et al. (2016) mention Vitrea transsylvanica (Mollusca: Gastropoda), Micrargus georgescuae (Chelicerata: Araneae), and Ligidium germanicum (Crustacea: Isopoda) as cold-adapted mountain species from the lower parts of a collapse doline in the Moravian Karst (Czech Republic). Similar results were published by Raschmanová et al. (2018), showing that collapse dolines in Slovakia harbour disproportionately high biodiversity and numbers of cold-adapted springtail species (Collembola). We also found that dolines may act as key habitats for a number of ants adapted to warmer and/or drier (e.g., Formica cunicularia, M. sabuleti, and Tapinoma subboreale) conditions. The main reason for this is that the south-facing slopes of dolines receive more insolation during daytime than the other microhabitats (Bárány-Kevei, 1999). This is in line

with Růžička et al. (2016) who found that south-facing slopes and edges of dolines may provide habitats for warm-adapted steppe species (e.g., *Danacea pallipes*, Insecta: Coleoptera). Our results suggest that dolines in Europe have the potential to buffer climate fluctuations. Because only a few areas have the potential to act as refugia with important ecological functions during both cooler and warmer climate periods (Harrison and Noss, 2017; Mokany et al., 2017), dolines may be particularly important for maintaining biodiversity through time (Bátori et al., 2017).

We found that the number of ant species associated primarily with forest habitats (e.g., F. truncorum, M. ruginodis, and L. platythorax) was the highest in grassland dolines, whereas only one species (M. ruginodis) with few occurrences was observed in forested sites. In Central Europe, Fagus sylvatica forests with lowresource availability are among the most species poor habitats, usually with only a few ant species and very low population densities (Dolek et al., 2009; Wiezik et al., 2015). In our previous study (Bátori et al., 2019), we detected a significant effect of slope exposure on ant species composition within the grassland dolines in the Bükk Mts (e.g., M. ruginodis dominated the north-facing slopes). This indicates that certain microhabitats in dolines have the potential to maintain suitable microclimate and resource availability for ant species associated primarily with forest habitats in landscapes where the dominant forest type is unfavourable for them. Considering that the study area in the Bükk Mts is dominated by Fagus sylvatica forests, these dolines may also function as stepping stones for some ant species, playing a crucial role for species' dispersal among distant habitats (cf. Saura et al., 2014). Similar patterns can also be observed in other island-like systems, such as under scattered trees in woodpastures (Tölgyesi et al., 2018) or on rocky outcrops and slopes in agricultural landscapes (Bellemare et al., 2002).

Generally, environmental heterogeneity has a positive effect on ant distribution (Ribas et al., 2003; Fagundes et al., 2015; Tavella et al., 2018), decreasing the monopolisation of space by behaviourally dominant species, thereby, providing more hiding places and foraging opportunities for subordinate species (Gibb and Parr, 2010; Koptur et al., 2010). Our results stand in contrast with these studies to some extent, indicating that baits in dolines were visited more frequently by moderately aggressive (e.g., M. ruginodis) and less by subordinate species (e.g., M. scabrinodis) than those on the plateau. Part of the explanation may be that despite the greater environmental heterogeneity, the restricted space within dolines favours species that are more aggressive and better competitors. However, the influence of some environmental factors, such as higher resource availability within dolines, may relax interspecific competition and can increase the probability of species coexistence (Ribas et al., 2003; Blüthgen et al., 2004).

Gauging the effects of surrounding habitat patterns on the species composition of island-like habitats and microrefugia is challenging because of the many possible interactions between environmental variables and species distributions (Deák et al., 2020; Gentili et al., 2020). For instance, the species composition of ant assemblages in habitat fragments is not only influenced by several abiotic and biotic factors but also by colonization–extinction dynamics and dispersal limitation (Crist, 2009). In

this study, we investigated how certain functional groups and traits in ant assemblages of dolines change with Fagus sylvatica cover outside the dolines. We found that the number of species adapted to intermediate moisture conditions was the highest in grassland dolines surrounded by large amounts of forests (type 3 dolines). These patterns are likely associated with the ecological effects of these forests (i.e., the edge effect) on the habitat structure, microclimate and resource availability of doline microhabitats. As key components of landscapes, edges have received considerable attention both in anthropogenic and natural ecosystems (Erdős et al., 2019), because they may control the flows of organisms (Wiens, 1992; Cadenasso et al., 2003), influence species interactions (Fagan et al., 1999), functional characteristics (Gallé et al., 2018) and evolutionary processes (Kevey and Borhidi, 1998), and serve as refuges for a wide range of organisms (Erdős et al., 2014). It can also be assumed that at least some environmental factors are intermediate at forest edges, providing habitats for many species adapted to intermediate environmental conditions (e.g., moisture or temperature), as we observed in this study (e.g., L. platythorax, M. lobicornis, and M. scabrinodis were relatively frequent in type 3 dolines). We also found that the number of aggressive species with mostly monogynous colonies was highest in type 3 dolines. One of the possible reasons for this phenomenon is that queens of monogynous colonies tend to disperse by flying to establish a new colony, inhabiting semi-isolated habitats more easily than queens of polygynous colonies that tend to disperse by budding or nest splitting (Czechowski et al., 2012; Seifert, 2018). Moreover, considering that Fagus sylvatica forests provide unfavourable habitats for ants (Dolek et al., 2009; Wiezik et al., 2015), some aggressive species associated with forest habitats (e.g., F. truncorum and L. platythorax) could find suitable habitats in these dolines (e.g., via fallen twigs and higher moss cover) for their foraging and nesting activities.

Global warming has been predicted to continue altering vegetation and landscape patterns in karst landscapes (Walther et al., 2002), likely resulting in significant changes in the composition of arthropod assemblages (Wise and Lensing, 2019; Prather et al., 2020). Based on their historical and current ability to facilitate the persistence of species (Bátori et al., 2017; Raschmanová et al., 2018), dolines in Europe will likely provide important safe havens, where different functional groups of species can survive for a longer period of time. However, changes to the biological environment (e.g., surrounding vegetation cover and forest height) caused by anthropogenic impacts (e.g., changes in forestry activity or in grazing/mowing regimes) may alter the species composition of dolines and their capacity to provide microrefugia for vulnerable species (cf. Liu et al., 2019; Bátori et al., 2020; Kermavnar et al., 2020). For instance, overgrazing is threatening the survival of vulnerable plants in the dolines of the Greek Archipelago (Egli et al., 1990). Therefore, careful conservation and management planning would be essential to maximise the resilience of karst landscapes to global warming. In the Bükk Mts, moderate grazing and/or mowing would be desirable in order to prevent shrub encroachment into the grassland microhabitats of dolines to ensure the optimal habitat and vegetation structure for ants. Preserving the current

distribution of forests and grassland patches may ensure that diverse ant assemblages are maintained by doline microhabitats.

# CONCLUSION

Our results demonstrate strong relationships between habitat heterogeneity and the distribution of functional groups and traits among ant assemblages in karst ecosystems. We found that dolines may harbour specific functional groups and traits that are rare in the surrounding habitats, and that vegetation structure on the surrounding plateau may have the potential to influence the functional diversity of ant assemblages within dolines. Species for which a given habitat becomes environmentally unsuitable may find shelter in doline microhabitats, but species can also use dolines as stepping stones during their range expansion. Our findings indicate that ants are reliable model organisms and possible indicators for identifying those locations that have the capacity to provide safe havens for different functional groups of species. Proper management and conservation of these safe havens may mitigate the rate of biodiversity loss under global warming.

# DATA AVAILABILITY STATEMENT

The original contributions presented in the study are included in the article/**Supplementary Material**, further inquiries can be directed to the corresponding author.

# **AUTHOR CONTRIBUTIONS**

ZB, GL, AV, and IM conceived and designed the experiment. GL, GM, OJ, DA, and IM conducted the field and lab work. CT performed the statistical analyses. ZB wrote the first draft of the manuscript. All authors contributed to content and editing of the manuscript.

# FUNDING

This research was funded by the NKFI K 124796 grant. The contribution of CT was supported by the NKFI PD 132131 grant. BD was supported by the NKFI KH 130338 and NKFI FK 135329 grants. OV was supported by the NKFI FK 124404 grant. GM was supported by the UNKP-19-3-SZTE-199 "New National Excellence Programme" of the Ministry of Human Capacities of Hungary. DA was supported by the Hungarian government through the Stipendium Hungaricum Scholarship Programme.

# SUPPLEMENTARY MATERIAL

The Supplementary Material for this article can be found online at: https://www.frontiersin.org/articles/10.3389/fevo.2020. 613738/full#supplementary-material

# REFERENCES

- Andersen, A. N. (1986). Diversity, seasonality and community organization of ants at adjacent heath and woodland sites in South-Eastern Australia. *Aust. J. Zool.* 34, 53–64. doi: 10.1071/zo9860053
- Andersen, A. N. (1997). Using ants as bioindicators: Multiscale issues in ant community ecology. *Conserv. Ecol.* 1:8. doi: 10.1046/j.1365-2664.2002. 00704.x
- Arnan, X., Cerdá, X., and Retana, J. (2014). Ant functional responses along environmental gradients. J. Anim. Ecol. 83, 1398–1408. doi: 10.1111/1365-2656. 12227
- Ashcroft, M. B. (2010). Identifying refugia from climate change. J. Biogeogr. 37, 1407-1413.
- Báldi, A. (2008). Habitat heterogeneity overrides the species-area relationship. J. Biogeogr. 35, 675–681. doi: 10.1111/j.1365-2699.2007. 01825.x
- Bárány-Kevei, I. (1999). Microclimate of karstic dolines. Acta Climatologica Universitatis Szegediensis 32–33, 19–27. doi: 10.3406/karst.2001. 2481
- Bates, D., Maechler, M., and Bolker, B. (2013). *lme4: Linear Mixedeffects Models Using S4 Classes. R Package Version 0.999999-2.* Available online at: http://cran. r-project.org/package=lme4. (accessed July 1, 2020).
- Bátori, Z., Csiky, J., Erdős, L., Morschhauser, T., Török, P., and Körmöczi, L. (2009). Vegetation of the dolines in Mecsek mountains (South Hungary) in relation to the local plant communities. *Acta Carsolog.* 38, 237–252. doi: 10. 3986/ac.v38i2-3.125
- Bátori, Z., Csiky, J., Farkas, T., Vojtkó, A. E., Erdős, L., Kovács, D., et al. (2014). The conservation value of karst dolines for vascular plants in woodland habitats of Hungary: refugia and climate change. *Int. J. Speleol.* 43, 15–26. doi: 10.5038/ 1827-806X.43.1.2
- Bátori, Z., Vojtkó, A., Farkas, T., Szabó, A., Havadtői, K., Vojtkó, A. E., et al. (2017). Large- and small-scale environmental factors drive distributions of cool-adapted plants in karstic microrefugia. *Ann. Bot. London* 119, 301–309. doi: 10.1093/aob/mcw233
- Bátori, Z., Vojtkó, A., Keppel, G., Tölgyesi, C., Čarni, A., Zorn, M., et al. (2020). Anthropogenic disturbances alter the conservation value of karst dolines. *Biodivers. Conserv.* 29, 503–525. doi: 10.1007/s10531-019-01896-4
- Bátori, Z., Vojtkó, A., Maák, I., Lőrinczi, G., Farkas, T., Kántor, N., et al. (2019). Karst dolines provide diverse microhabitats for different functional groups in multiple phyla. *Sci. Rep. UK* 9:7176. doi: 10.1038/s41598-019-43603-x
- Battisti, C., Giardini, M., Marini, F., Di Rocco, L., Dodaro, G., and Vignoli, L. (2017). Diversity metrics, species turnovers and nestedness of bird assemblages in a deep karst sinkhole. *Isr. J. Ecol. Evol.* 63, 8–16. doi: 10.1163/22244662-06301009
- Beck von Mannagetta, G. (1906). Die umkehrung der pflanzenregionen in den dolinen des karstes. Sitzungsberichte der Kaiserliche Akademie der Wissenschaften in Wien 65, 3–4. doi: 10.1515/9783111410272-001
- Bellemare, J., Motzkin, G., and Foster, D. R. (2002). Legacies of the agricultural past in the forested present: an assessment of historical land-use effects on rich mesic forests. J. Biogeogr. 29, 1401–1420. doi: 10.1046/j.1365-2699.2002. 00762.x
- Blüthgen, N., Stork, N. E., and Fiedler, K. (2004). Bottom-up control and co-occurrence in complex communities: honeydew and nectar determine a rainforest ant mosaic. *Oikos* 106, 344–358. doi: 10.1111/j.0030-1299.2004. 12687.x
- Bolton, B. (2020). An Online Catalog of the Ants of the World. Available online at: http://antcat.org (accessed June 2, 2020).
- Cadenasso, M. L., Pickett, S. T. A., Weathers, K. C., Bell, S. S., Benning, T. L., Carreiro, M. M., et al. (2003). An interdisciplinary and synthetic approach to ecological boundaries. *BioScience* 53, 717–722. doi: 10.1641/0006-3568(2003) 053[0717:aiasat]2.0.co;2
- Crist, T. O. (2009). Biodiversity, species interactions, and functional role of ants (*Hymenoptera: Formicidae*) in fragmented landscapes: a review. *Myrmecol.* News 12, 3–13.
- Culver, D. C. (2000). Hotspots of subterranean biodiversity in caves and wells. *J. Cave Karst Stud.* 62, 11–17.

- Czechowski, W., Radchenko, A., Czechowska, W., and Vepsäläinen, K. (2012). *The Ants (Hymenoptera, Formicidae) of Poland with Reference to the Myrmecofauna of Europe.* Warszawa: MIZ PAS.
- de Aguiar-Campos, N., Maia, V. A., da Silva, W. B., de Souza, C. R., and dos Santos, R. M. (2020). Can fine-scale habitats of limestone outcrops be considered litho-refugia for dry forest tree lineages? *Biodivers. Conserv.* 29, 1009–1026. doi: 10.1007/s10531-019-01923-4
- Deák, B., Valkó, O., Nagy, D. D., Török, P., Torma, A., Lőrinczi, G., et al. (2020). Habitat islands outside nature reserves - Threatened biodiversity hotspots of grassland specialist plant and arthropod species. *Biol. Conserv.* 241:108254. doi: 10.1016/j.biocon.2019.108254
- Dobrowski, S. Z. (2010). A climatic basis for microrefugia: the influence of terrain on climate. *Glob. Change Biol.* 17, 1022–1035. doi: 10.1111/j.1365-2486.2010. 02263.x
- Dolek, M., Freese-Hager, A., Bussler, H., Floren, A., Liegl, A., and Schmidl, J. (2009). Ants on oaks: effects of forest structure on species composition. J. Insect. Conserv. 13, 367–375. doi: 10.1007/s10841-008-9181-2
- Dövényi, Z. (ed.) (2010). *Magyarország Kistájainak Katasztere*. Budapest: MTA Földrajztudományi Kutatóintézet.
- Egli, B. R., Gerstberger, P., Greuter, W., and Risse, H. (1990). *Horstrissea dolinicola*, a new genus and species of umbels (*Umbelliferae*, *Apiaceae*) from Kriti (Greece). *Willdenowia* 19, 389–399.
- Eldridge, D. J., Oliver, I., Valc, J., Travers, S. K., and Delgado-Baquerizo, M. (2020). Grazing and aridity have contrasting effects on the functional and taxonomic diversity of ants. *Basic. Appl. Ecol.* 48, 73–82. doi: 10.1016/j.baae.2020.07.003
- Erdős, L., Krstonošić, D., Kiss, P. J., Bátori, Z., Tölgyesi, C., and Škvorc, Ž (2019). Plant composition and diversity at edges in a semi-natural forest-grassland mosaic. *Plant Ecol.* 220, 279–292. doi: 10.1007/s11258-019-00913-4
- Erdős, L., Tölgyesi, C., Horzse, M., Tolnay, D., Hurton, Á, Schulcz, N., et al. (2014). Habitat complexity of the Pannonian forest-steppe zone and its nature conservation implications. *Ecol. Complex* 17, 107–118. doi: 10.1016/j.ecocom. 2013.11.004
- Fagan, W. F., Cantrell, R. S., and Cosner, C. (1999). How habitat edges change species interactions. Am. Nat. 153, 165–182. doi: 10.2307/24 63579
- Fagundes, R., Anjos, D. V., Carvalho, R., and Del-Claro, K. (2015). Availability of food and nesting-sites as regulatory mechanisms for the recovery of ant diversity after fire disturbance. *Sociobiology* 62, 1–9. doi: 10.13102/sociobiology. v62i1.1-9
- Fayle, T. M., Turner, E. C., and Foster, W. A. (2013). Ant mosaics occur in SE Asian oil palm plantation but not rain forest and are influenced by the presence of nest-sites and non-native species. *Ecography* 36, 1051–1057. doi: 10.1111/j. 1600-0587.2012.00192.x
- Folgarait, P. J. (1998). Ant biodiversity and its relationship to ecosystem functioning: a review. *Biodivers. Conserv.* 7, 1221–1244. doi: 10.1023/A: 1008891901953
- Fox, J., and Weisberg, S. (2011). An {R} Companion to Applied Regression, Second Edn. Thousand Oaks CA: Sage.
- Gallé, L. (2017). Climate change impoverishes and homogenizes ants' community structure: a long term study. *Community Ecol.* 18, 128–136. doi: 10.1556/168. 2017.18.2.2
- Gallé, R., Szabó, Á, Császár, P., and Torma, A. (2018). Spider assemblage structure and functional diversity patterns of natural forest steppes and exotic forest plantations. *Forest Ecol. Manag.* 411, 234–239. doi: 10.1016/j.foreco.2018. 01.040
- Gentili, R., Baroni, C., Caccianiga, M., Armiraglio, S., Ghiani, A., and Citterio, S. (2015). Potential warm-stage microrefugia for alpine plants: Feedback between geomorphological and biological processes. *Ecol. Complex* 21, 87–99. doi: 10. 1016/j.ecocom.2014.11.006
- Gentili, R., Baroni, C., Panigada, C., Rossini, M., Tagliabue, G., Armiraglio, S., et al. (2020). Glacier shrinkage and slope processes create habitat at high elevation and microrefugia across treeline for alpine plants during warm stages. *Catena* 193:104626. doi: 10.1016/j.catena.2020.104626
- Gibb, H., and Parr, C. L. (2010). How does habitat complexity affect ant foraging success? a test using functional measures on three continents. *Oecologia* 164, 1061–1073. doi: 10.1007/s00442-010-1703-4
- Guilherme, D. R., Souza, J. L. P., Franklin, E., Pequeno, P. A. C. L., das Chagas, A. C., and Baccaro, F. B. (2019). Can environmental complexity predict

functional trait composition of ground-dwelling ant assemblages? a test across the Amazon Basin. *Acta Oecol.* 99:103434. doi: 10.1016/j.actao.2019.05.004

- Harrison, S., and Noss, R. (2017). Endemism hotspots are linked to stable climatic refugia. Ann. Bot. London 119, 207–214. doi: 10.1093/aob/ mcw248
- Heinze, J. (2017). Life-history evolution in ants: the case of Cardiocondyla. P. Roy. Soc. B Biol. Sci. 284:20161406. doi: 10.1098/rspb.2016.1406
- Hortal, J., Triantis, K. A., Meiri, S., Thébault, E., and Sfenthourakis, S. (2009). Island species richness increases with habitat diversity. Am. Nat. 174, E205–E217.
- Horvat, I. (1953). Vegetacija ponikava. Geografski Glasnik 14-15, 1-25. doi: 10. 15291/geoadria.23
- Johnson, N. F. (2007). Hymenoptera Name Server (version 1.5). Available online at: http://osuc.biosci.ohio-state.edu/ (accessed June 2, 2020).
- Kemencei, Z., Farkas, R., Páll-Gergely, B., Vilisics, F., Nagy, A., Hornung, E., et al. (2014). Microhabitat associations of land snails in forested dolinas: implications for coarse filter conservation. *Community Ecol.* 15, 180–186. doi: 10.1556/ comec.15.2014.2.6
- Keppel, G., Mokany, K., Wardell-Johnson, G. W., Phillips, B. L., Welbergen, J. A., and Reside, A. E. (2015). The capacity of refugia for conservation planning under climate change. *Front. Ecol. Environ.* 13:106–112. doi: 10.1890/140055
- Keppel, G., Ottaviani, G., Harrison, S., Wardell-Johnson, G. W., Marcantonio, M., and Mucina, L. (2018). Towards an eco-evolutionary understanding of endemism hotspots and refugia. *Ann. Bot. London* 122, 927–934. doi: 10.1093/ aob/mcy173
- Keppel, G., Van Niel, K. P., Wardell-Johnson, G. W., Yates, C. J., Byrneet, M., Mucina, L., et al. (2012). Refugia: identifying and understanding safe havens for biodiversity under climate change. *Global Ecol. Biogeogr.* 21, 393–404. doi: 10.1111/j.1466-8238.2011.00686.x
- Kermavnar, J., Ferlan, M., Marinšek, A., Eler, K., Kobler, A., and Kutnar, L. (2020). Effects of various cutting treatments and topographic factors on microclimatic conditions in Dinaric fir-beech forests. *Agr. Forest Meteorol.* 295:108186. doi: 10.1016/j.agrformet.2020.108186
- Kevey, B., and Borhidi, A. (1998). Top-forest (Aconito anthorae-Fraxinetum orni), a special ecotonal case in the phytosociological system (Mecsek Mts., South Hungary). Acta Bot. Hung. 41, 27–121.
- Kobal, M., Bertoncelj, I., Pirotti, F., Dakskobler, I., and Kutnar, L. (2015). Using lidar data to analyse sinkhole characteristics relevant for understory vegetation under forest cover – case study of a high karst area in the Dinaric mountains. *PLoS One* 10:e0122070. doi: 10.1371/journal.pone.0122070
- Koptur, S., William, P., and Olive, Z. (2010). Ants and plants with extrafloral nectaries in fire successional habitats on Andros (Bahamas). *Florida Entomol.* 93, 89–99. doi: 10.1653/024.093.0112
- Lehner, M., Whiteman, C. D., and Dorninger, M. (2017). Inversion build-up and cold-air outflow in a small alpine sinkhole. *Bound-Lay. Meteorol.* 163, 497–522. doi: 10.1007/s10546-017-0232-7
- Liu, R., Zhang, Z., Shen, J., and Wang, Z. (2019). Bryophyte diversity in karst sinkholes affected by different degrees of human disturbance. *Acta Soc. Bot. Pol.* 88:3620. doi: 10.5586/asbp.3620
- Mammola, S., Aharon, S., Seifan, M., Lubin, Y., and Gavish-Regev, E. (2019). Exploring the interplay between local and regional drivers of distribution of a subterranean organism. *Diversity* 11:119. doi: 10.3390/d11080119
- Michalak, J. L., Stralberg, D., Cartwright, J. M., and Lawler, J. J. (2020). Combining physical and species-based approaches improves refugia identification. *Front. Ecol. Environ.* 18, 254–260. doi: 10.1002/fee.2207
- Mokany, K., Jordan, G. J., Harwood, T. D., Harrison, P. A., Keppel, G., Gilfedder, L., et al. (2017). Past, present and future refugia for Tasmania's palaeoendemic flora. J. Biogeogr. 44, 1537–1546. doi: 10.1111/jbi.12927
- Ottaviani, G., Keppel, G., Götzenberger, L., Harrison, S., Opedal, Ø. H., Conti, L., et al. (2020). Linking plant functional ecology to island biogeography. *Trends Plant Sci.* 25, 329–339. doi: 10.1016/j.tplants.2019.12.022
- Ottonetti, L., Tucci, L., and Santini, G. (2006). Recolonization patterns of ants in a rehabilitated lignite mine in central Italy: potential for the use of Mediterranean ants as indicators of restoration processes. *Restor. Ecol.* 14, 60–66. doi: 10.1111/j.1526-100x.2006.00105.x
- Özkan, K., Gulsoy, S., Mert, A., Ozturk, M., and Muys, B. (2010). Plant distribution-altitude and landform relationships in karstic sinkholes of Mediterranean region of Turkey. J. Environ. Biol. 31, 51–61.

- Pacheco, R., and Vasconcelos, H. L. (2012). Habitat diversity enhances ant diversity in a naturally heterogeneous Brazilian landscape. *Biodivers. Conserv.* 21, 797– 809. doi: 10.1007/s10531-011-0221-y
- Prather, R. M., Castillioni, K., Welti, E. A. R., Kaspari, M., and Souza, L. (2020). Abiotic factors and plant biomass, not plant diversity, strongly shape grassland arthropods under drought conditions. *Ecology* 101:e03033. doi: 10.1002/ecy. 3033
- R Development Core Team. (2017). R: a Language and Environment for Statistical Computing. Vienna: R Foundation for Statistical Computing.
- Raschmanová, N., Miklisová, D., and Kováč, L. (2013). Soil Collembola communities along a steep microclimatic gradient in the collapse doline of the Silická l'adnica Cave, Slovak Karst (Slovakia). *Biologia* 68, 470–478.
- Raschmanová, N., Miklisová, D., and Kováč, Ľ. (2018). A unique small-scale microclimatic gradient in a temperate karst harbours exceptionally high diversity of soil Collembola. *Int. J. Speleol.* 47, 247–262. doi: 10.5038/1827-806X. 47.2.2194
- Raschmanová, N., Miklisová, D., Kováč, Ľ., and Šustr, V. (2015). Community composition and cold tolerance of soil Collembola in a collapse karst doline with strong microclimate inversion. *Biologia* 70, 802–811. doi: 10.1515/biolog-2015-0095
- Ribas, C. R., Schoereder, J. H., Pic, M., and Soares, S. M. (2003). Tree heterogeneity, resource availability, and larger scale processes regulating arboreal ant species richness. *Austral Ecol.* 28, 305–314. doi: 10.1046/j.1442-9993.2003.01290.x
- Růžička, V., Mlejnek, R., Juřičková, L., Tajovský, K., Šmilauer, P., and Zajíček, P. (2016). Invertebrates of the Macocha Abyss (*Moravian Karst*, Czech Republic). *Acta Carsolog.* 45, 71–84. doi: 10.3986/ac.v45i1.896
- Sanders, N. J. (2002). Elevational gradients in ant species richness: area, geometry, and Rapoport's rule. *Ecography* 25, 25–32. doi: 10.1034/j.1600-0587.2002. 250104.x
- Saura, S., Bodin, Ö, and Fortin, M.-J. (2014). Stepping stones are crucial for species' long-distance dispersal and range expansion through habitat networks. J. Appl. Ecol. 51, 171–182. doi: 10.1111/1365-2664.12179
- Seifert, B. (2018). *The Ants of Central and North Europe*. Tauer: Lutra Verlags- and Vertriebsgesellschaft.
- Stewart, J. R., Lister, A. M., Barnes, I., and Dalén, L. (2010). Refugia revisited: individualistic responses of species in space and time. P. Roy. S. B Biol. Sci. 277, 661–671. doi: 10.1098/rspb.2009.1272
- Su, Y., Tang, Q., Mo, F., and Xue, Y. (2017). Karst tiankengs as refugia for indigenous tree flora amidst a degraded landscape in southwestern China. Sci Rep. UK 7:4249. doi: 10.1038/s41598-017-04592-x
- Tavella, J., Alvarez Pringles, A. P., and Cagnolo, L. (2018). Determinants of ant species spatial distribution in habitats from central Argentina. *Community Ecol.* 19, 300–310. doi: 10.1556/168.2018.19.3.11
- Tichý, L. (2002). JUICE, software for vegetation classification. J. Veg. Sci. 13, 451–453. doi: 10.1111/j.1654-1103.2002.tb02069.x
- Tichý, L., and Chytrý, M. (2006). Statistical determination of diagnostic species for site groups of unequal size. J. Veg. Sci. 17, 809–818. doi: 10.1111/j.1654-1103. 2006.tb02504.x
- Tölgyesi, C., Bátori, Z., Gallé, R., Urák, I., and Hartel, T. (2018). Shrub encroachment under the trees diversifies the herb layer in a Romanian silvopastoral system. *Rangel. Ecol. Manag.* 71, 571–577. doi: 10.1016/j.rama. 2017.09.004
- Vilisics, F., Sólymos, P., Nagy, A., Farkas, R., Kemencei, Z., and Hornung, E. (2011). Small scale gradient effects on isopods (*Crustacea: Oniscidea*) in karstic sinkholes. *Biologia* 66, 499–505. doi: 10.2478/s11756-011-0042-1
- Vojtkó, A. (2001). A Bükk Hegység Flórája. Eger: Sorbus 2001 Kiadó.
- Walther, G.-R., Post, E., Convey, P., Menzel, A., Parmesan, C., Beebee, T. J. C., et al. (2002). Ecological responses to recent climate change. *Nature* 416, 389–437. doi: 10.1038/416389a
- White, W. B., Culver, D. C., Herman, J. S., Kane, T. C., and Mylroie, J. E. (1995). Karst Lands. The dissolution of carbonate rock produces unique landscapes and poses significant hydrological and environmental concerns. *Am. Sci.* 83, 450–459.
- Whiteman, C. D., Haiden, T., Pospichalm, B., Eisenbach, S., and Steinacker, R. (2004). Minimum temperatures, diurnal temperature ranges, and temperature inversion in limestone sinkholes of different sizes and shapes. J. Appl. Meteorol. 43, 1224–1236. doi: 10.1175/1520-0450(2004)043<1224:mtdtra>2.0.co;2

Wiens, J. A. (1992). "Ecological flows across landscape boundaries: a conceptual overview," in Landscape Boundaries: Consequences for Biotic Diversity and Ecological Flows, eds J. Hansen and F. di Castri (New York: Springer-Verlag), 217–235.

Wiezik, M., Svitok, M., Wieziková, A., and Dovčiak, M. (2015). Identifying shifts in leaf-litter ant assemblages (*Hymenoptera: Formicidae*) across ecosystem boundaries using multiple sampling methods. *PLoS One* 10:e0134502. doi: 10. 1371/journal.pone.0134502

Wise, D. H., and Lensing, J. R. (2019). Impacts of rainfall extremes predicted by climate-change models on major trophic groups in the leaf litter arthropod community. J. Anim. Ecol. 88, 1486–1497. doi: 10.1111/1365-2656.13046 **Conflict of Interest:** The authors declare that the research was conducted in the absence of any commercial or financial relationships that could be construed as a potential conflict of interest.

Copyright © 2020 Bátori, Lőrinczi, Tölgyesi, Módra, Juhász, Aguilon, Vojtkó, Valkó, Deák, Erdős and Maák. This is an open-access article distributed under the terms of the Creative Commons Attribution License (CC BY). The use, distribution or reproduction in other forums is permitted, provided the original author(s) and the copyright owner(s) are credited and that the original publication in this journal is cited, in accordance with accepted academic practice. No use, distribution or reproduction is permitted which does not comply with these terms.