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12 **Qualitative discrimination of coal aerosols**
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14 **by using the statistical evaluation of**
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16 **laser-induced breakdown spectroscopy data**
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21 D.J. Palásti^{1,2}, A. Metzinger¹, T. Ajtai³, Z. Bozóki^{3,4}, B. Hopp^{2,3}, É. Kovács-Széles⁵, G. Galbács^{1,2*}
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25 ¹Dept. of Inorganic and Analytical Chemistry, University of Szeged,
26 Dóm sq. 7, 6720 Szeged, Hungary
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30 ²Dept. of Materials Science, Interdisciplinary Excellence Centre, University of Szeged,
31 6720 Szeged, Dugonics sq. 13, Hungary
32
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36 ³Dept. of Optics and Quantum Electronics,
37 University of Szeged, Dóm sq. 9, 6720 Szeged, Hungary
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41 ⁴MTA-SZTE Research Group on Photoacoustic Spectroscopy, Dóm sq. 9, 6720 Szeged, Hungary
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45 ⁵MTA Centre for Energy Research, Konkoly-Thege M. str. 29-33, 1121 Budapest, Hungary
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55 *Corresponding author. E-mail: galbx@chem.u-szeged.hu
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62 **Abstract:** The present work demonstrates the capabilities of laser-induced breakdown
63 spectroscopy (LIBS) in the qualitative analysis of carbonaceous aerosols. The model aerosols
64 studied were generated by laser ablation in a nitrogen flow from five commercial coal samples
65 (lignite, anthracite, Pécs-vasas brown coal, Polish brown coal, Czech brown coal) and contained
66 sub-micron particles in a concentration exceeding 10^6 cm^{-3} . Features of the LIBS spectra of these
67 aerosol samples were characterized and it is showed that the particle detection frequency
68 (expressed as the number of particle hits referenced to the total number of laser shots
69 delivered) correlates with the mass concentration of the aerosol. The detection limit for coal
70 aerosols was also estimated and found to be about $600 \text{ pg}\cdot\text{mm}^{-3}$, meaning that for detectability
71 with the present experimental system, either the diameter of individual particles should be over
72 $2.3 \text{ }\mu\text{m}$ or their number concentration has to be large enough to exceed the above mass
73 concentration detection limit.
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80 The possibilities for coal classification based on the statistical evaluation of the LIBS
81 spectra of their aerosols was also investigated in detail in the laboratory. Simple comparative
82 functions (overlapping integrals, sum of squared deviations, linear correlation) were found not
83 to be efficient in the discrimination, even when facilitated by spectral masking, but multivariate
84 methods (classification tree, linear and quadratic discrimination analysis) gave significantly
85 better results. The effect of data normalization and data compression (by the multivariate curve
86 resolution alternating least squares methodology) prior to modeling was also tested and it was
87 found that their influence on the classification accuracy is not always positive, if any. The best
88 performance was showed by the classification tree method (without data compression), which
89 had a good overall accuracy of 87.2%. The validation of the model was assessed by calculating
90 the repeatability of the classification accuracy from 15 repetitions using randomly selected
91 subsets of the spectra. This calculation gave a repeatability value of 2.2%, which shows that the
92 model is quite robust.
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100 **Keywords:** aerosols; coal; laser ablation; LIBS; multivariate statistics
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121 **1. INTRODUCTION**
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123 Carbonaceous particulate matter (CPM) is one of the most important member of anthropogenic
124 aerosols, and hence it is the topic of a gradually increasing scientific interest in environmental
125 science. On the global scale, CPM is targeted because it is a major contributor to global warming
126 due to its strong light absorption, while on the local scale it is an air pollutant with multiple
127 adverse health effects. From both aspects, the identification and source apportionment of CPM in
128 ambient aerosols is of essential importance [1-4].
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132 Earlier studies have already demonstrated that residential domestic coal burning alone
133 accounts for nearly half of the total CPM emission from fossil fuel combustion [5]. In spite of its
134 significant contribution to both the global radiative budget and also to the quality of air,
135 available data on its quantity and its emission characteristics is limited, partly because of the
136 negligible contribution of this type of fuels to the emission inventory of developed countries [5].
137 Moreover, related studies often only focus on mass emissions and rarely address the topic of
138 selective source identification.
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143 In fact, there is generally no widely accepted methodology for the selective and real-time
144 apportionment of CPM. Conventional approaches are off-line and are based on the collection of
145 aerosol followed by wet chemical or thermochemical analysis [6-8]. The off-line nature, the
146 moderate selectivity and integration time strongly limits the applicability of these approaches.
147 The recently appeared alternatives applicable to the real-time (in-situ) identification of ambient
148 CPM, such as aerosol mass spectrometry (AMS) [9], multi-wavelength photo-acoustic
149 spectroscopy (PAS) [10, 11] and laser-induced breakdown spectroscopy (LIBS) [12, 13] open up
150 new perspectives in this field. Considering the price and bulkiness of AMS instrumentation as
151 well as the indirect data provided by PAS, LIBS is the most promising out of the mentioned three
152 methodologies.
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157 LIBS is a sensitive, robust and fast laser atomic emission spectroscopy method, which is
158 based on the generation of an analytical microplasma on the surface of solid samples by using a
159 focused, high power laser beam. Portable and stand-off LIBS instrumentation is also available
160 commercially, thus this technique is very suitable for field measurements in many scenarios
161 including industrial, environmental or safety-related analytical tasks [14, 15]. Industrial
162 applications also include coal and fly ash analysis [16-18], as well as aerosol analysis in general.
163 In the latter area, it was shown that LIBS can be employed both in on-line (in situ, on an aerosol
164 beam) or off-line (e.g. following aerosol collection on a filter) modes. Comprehensive reviews on
165 the status of research on LIBS aerosol analysis were published by Hahn et al. [15, 19, 20].
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171 In recent years, some studies also successfully demonstrated the use of LIBS for the
172 analysis of coal particles. These studies focused either on the detection of particles or on the
173 quantitative analysis of some components, typically carbon, in the particles. For example, Stipe
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180 et al. quantified quartz in coal dust samples collected on filter media [21], in the off-line mode.
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182 The carbon content of coal aerosol and pulverized coal particles were successfully determined in
183 the on-line mode in several studies [13, 18, 22, 23]. Yao et al. designed an advanced, particle flow
184 focusing device for the efficient detection of coal particles [24]. Mukherjee and Cheng
185 investigated how different mass loadings in the plasma induce matrix effects during the
186 quantitative analysis of carbonaceous aerosols [12]. At the same time, no studies have addressed
187 the on-line, qualitative discrimination of coal aerosols according to their source coal type so far,
188 although the potential of LIBS for the classification/identification of samples has already been
189 successfully demonstrated on various samples, for example in microbiological [25-27], medical
190 [28, 29], industrial [16, 30, 31] or forensic [32-34] applications.

195 In the present study, we assessed the performance of LIBS combined with data
196 evaluation using multivariate statistics for the discrimination of different types of coal aerosols
197 in the laboratory. This is the first study in the literature, according to the best knowledge of the
198 authors, which attempts to use LIBS for the source coal type determination based on an on-line
199 coal aerosol analysis. Our model aerosols were generated from residential coal samples using
200 our recently introduced laser ablation methodology [35, 36] in nitrogen gas, in order to control
201 the properties of the aerosol particles and to reasonably mimic exhaust flue gases from air-
202 based coal combustion.
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208 **2. EXPERIMENTALS**

210 **2.1. Coal samples and aerosol generation**

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212 The set of samples used to generate coal aerosol consisted of five coals, including low-rank and
213 high-rank coals as well. The coal samples were obtained from commercial sources in Hungary.
214 These were identified as follows: lignite (LI), anthracite (AN), Pécs-vasas brown coal (PV), Polish
215 brown coal (PB), Czech brown coal (CB). Disk shaped target materials, with a diameter of 25 mm
216 and a thickness of about 10 mm, were fabricated from the irregularly shaped original coal
217 pieces.
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221 Aerosol samples were generated by using laser ablation. The experimental arrangement
222 and the formation mechanism of coal particles by laser ablation are described in details e.g. in
223 our former publications [35, 36]. The samples (targets) were implemented onto a rotating
224 sample holder in this ablation chamber and the focused laser beam of a KrF excimer laser was
225 directed onto the sample surface. The laser fluence, repetition rate (2.5 J/cm² and 10 Hz,
226 respectively) as well as the sample rotation speed were fixed to provide a stable particle number
227 concentration of about 5·10⁶ cm⁻³ for each sample. The generated aerosol plume was dispersed
228 in a high purity nitrogen gas (99.995% purity, Messer Hungarogáz Ltd.). Nitrogen gas was
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chosen in order to control the ablation environment and therefore the properties of the produced aerosol particles and partially also to reasonably mimic exhaust flue gases from air-based coal combustion which is mainly composed of nitrogen. The flow rate of the gas was controlled by a mass flow controller (MFC, Tylan 2900FC) and set at a value of 0.5 liter per minute. The coal aerosol stream was directed into a specially designed LIBS ablation chamber described in the next section.

Aerosols leaving the generation chamber were characterized by a scanning mobility particle sizer (SMPS), consisting of a Vienna-type differential mobility analyzer and a condensation particle counter (Grimm Aerosol Technik GmbH & Co.). Relevant properties of the generated coal aerosols can be seen in Table 1.

Coal sample (identifier)	Concentration (cm ⁻³)	Electrical Aerodynamic mobility particle diameter (μm)	Size distribution curve, FWHM (μm)
Lignite (LI)	8.87 × 10 ⁶	0.476	0.604
Anthracite (AN)	1.68 × 10 ⁶	0.527	0.684
Pécs-vasas brown coal (PV)	5.19 × 10 ⁶	0.238	0.308
Polish brown coal (PB)	7.82 × 10 ⁶	0.385	0.489
Czech brown coal (CB)	8.70 × 10 ⁶	0.465	0.566

Table 1. Characteristics of the coal aerosols generated by laser ablation for the purposes of the study.

2.2. LIBS instrumentation

The experiments were carried out in a flow-through, small volume measurement chamber constructed in-house for LIBS aerosol measurements. The aerosol flow from the generation chamber was directed through this second chamber and a Nd:YAG laser beam was focused into the chamber from above through a UV-grade fused silica laser window for LIB plasma generation inside the particle beam. The repetition rate of the plasma generating laser was set to a low value (ca. 0.3 Hz) thereby ensuring that the content of the chamber is completely renewed by the gas flow between laser shots. The Nd:YAG laser was operated in the single-pulse mode, emitting 50 mJ pulses of 4 ns duration at the 1064 nm fundamental wavelength (LIBScan 25+, Applied Photonics, UK). The fractional residual part of the laser beam was coupled out from the chamber through a laser mirror and a second window and led into a beam dump, in order to eliminate laser beam reflections inside the measurement chamber. The LIBS plasma was observed via two fused silica collimating lenses (COL-UV/VIS, Avantes, NL) implemented in two ports located on the sides of the chamber (looking onto the same spot in the chamber in the

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298 horizontal direction, with 90° angle between the optical axes of the two lenses). The collected
299 light was coupled into a two-channel fiber-optic CCD spectrometer (AvaSpec-FT2048, Avantes,
300 NL) using solarization-resistant, 200 μm diameter, 0.22 NA optical fibers (FCB-UV200-2-SR,
301 Avantes, NL). The measurement chamber was mounted on a translation stage (LT-1, Thorlabs
302 Inc., USA), which allowed bringing the laser focal spot inside the chamber in front of the light
303 collection lenses.
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307 Gating of the spectral data collection was achieved by the internal electronics of the
308 spectrometer, which was triggered by the laser power supply unit and continuously monitored
309 on a digital storage oscilloscope (TDS1002, Tektronix, USA). The minimum possible gate delay of
310 1 μs and gate width of 2 ms was set at the spectrometer. The double-channel spectrometer
311 allowed the recording of the plasma emission in the 198-318 nm and 344-888 nm spectral
312 ranges, with resolutions of 0.09 nm and 0.4 nm, respectively.
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317 **2.3. Data evaluation**

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319 Spectral line assignment in LIBS spectra was done using the NIST Atomic Spectra Database.
320 Spectrum normalization, by using the total integrated spectral intensity as reference value, was
321 also used during the data evaluation. Multivariate statistical data evaluation was performed in
322 the open source R Studio program environment, using the following software packages: MASS,
323 ALS, Chemometrics and RPART. Additional statistical evaluation and data plots were done in
324 Origin Pro 8.6 (OriginLab, USA). Calculations using comparative statistical functions were
325 executed in MS Office Excel 2010 (Microsoft, USA), by using Visual Basic for Applications macros.
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331 The statistical functions and methods used in the present work are known in the
332 literature and we also described them in detail in our earlier qualitative discrimination studies
333 performed on other sample types (e.g. [16, 31, 34]), so here we only briefly outline their concept.
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335 Three common comparative functions, namely *linear correlation* (LC), *sum of squared*
336 *deviations* (SSD) and *overlapping integral* (OI) were used in the simpler statistical calculations.
337 In our implementation, we normalize these functions to the range of 0 to 1. This range defines a
338 degree of similarity (Q), where a value of 1 means “full similarity” and $Q= 0$ indicates “full
339 dissimilarity” of the compared spectra [16, 31, 34].
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342 *Multivariate curve resolution - alternating least squares* (MCR-ALS) decomposes a 2D
343 dataset into the linear combination of component datasets [37, 38]. Generally, the process can be
344 described by the equation $D= C \times ST + E$, where D is the spectrum of the (composite) sample, E is
345 the error matrix, and ST and C are the column and row vectors of the reference spectra and
346 concentrations of the pure components, respectively. In the present application, MCR-ALS
347 decomposition was not used as a standalone evaluation method, but served as a "data
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357 compression" step, which generated a set of virtual components (ST and C), constituting a
358 significantly smaller data matrix than the original set of sample spectra.
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360 *Classification tree* (CLT) is a method that produces a list of conditions, based on which
361 the features of a dataset (sample spectra) can be classified, or grouped. In the case of spectral
362 datasets, these conditions refer to the intensity and positional relations of peaks in the spectra.
363 The series of these conditions make up the evaluational construct, which can be best
364 represented as a tree structure [39, 40].
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367 *Discriminant analysis* (DA) finds a set of prediction equations based on independent
368 variables that are used to classify individuals into groups [41]. The methodology used to
369 complete a discriminant analysis is similar to regression analysis. Discrimination is achieved by
370 setting the variate's weight for each variable to maximize the between-the-groups variance
371 relative to the within-the-group variance. DA is most often used either in a linear or quadratic
372 form. The so called linear DA (LDA) assumes that the observations within each groups are
373 drawn from a multivariate Gaussian distribution and the covariance of the predictor variables
374 are common across all levels of the response variable. Quadratic DA (QDA) also assumes
375 normality of the data, but allows each class to have its own covariance matrix. The result is that
376 QDA is usually more flexible. LDA and QDA require the number of predictor variables to be less
377 than the sample size.
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385 **3. RESULTS AND DISCUSSION**

386 **3.1. Spectrum characterization**

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389 500 spectra for each of the aerosol samples were collected in nitrogen gas. Only Vis-range data
390 were included in further investigations as they were more eventful. The typical average spectra
391 of the aerosol samples, as well as the assignation of the most prominent spectral lines, are
392 shown in Fig. 1. Among the most prominent observed lines in this range were CN bands, Ca and
393 N peaks – other spectral features had small intensities. It is expected that the spectral intensities
394 observed also depend on the water and volatile matter content of coal samples [42-44].
395 Nevertheless, reasonably good signal-to-noise spectra could be recorded for all samples during
396 the ca. 25 minutes long measurement sessions.
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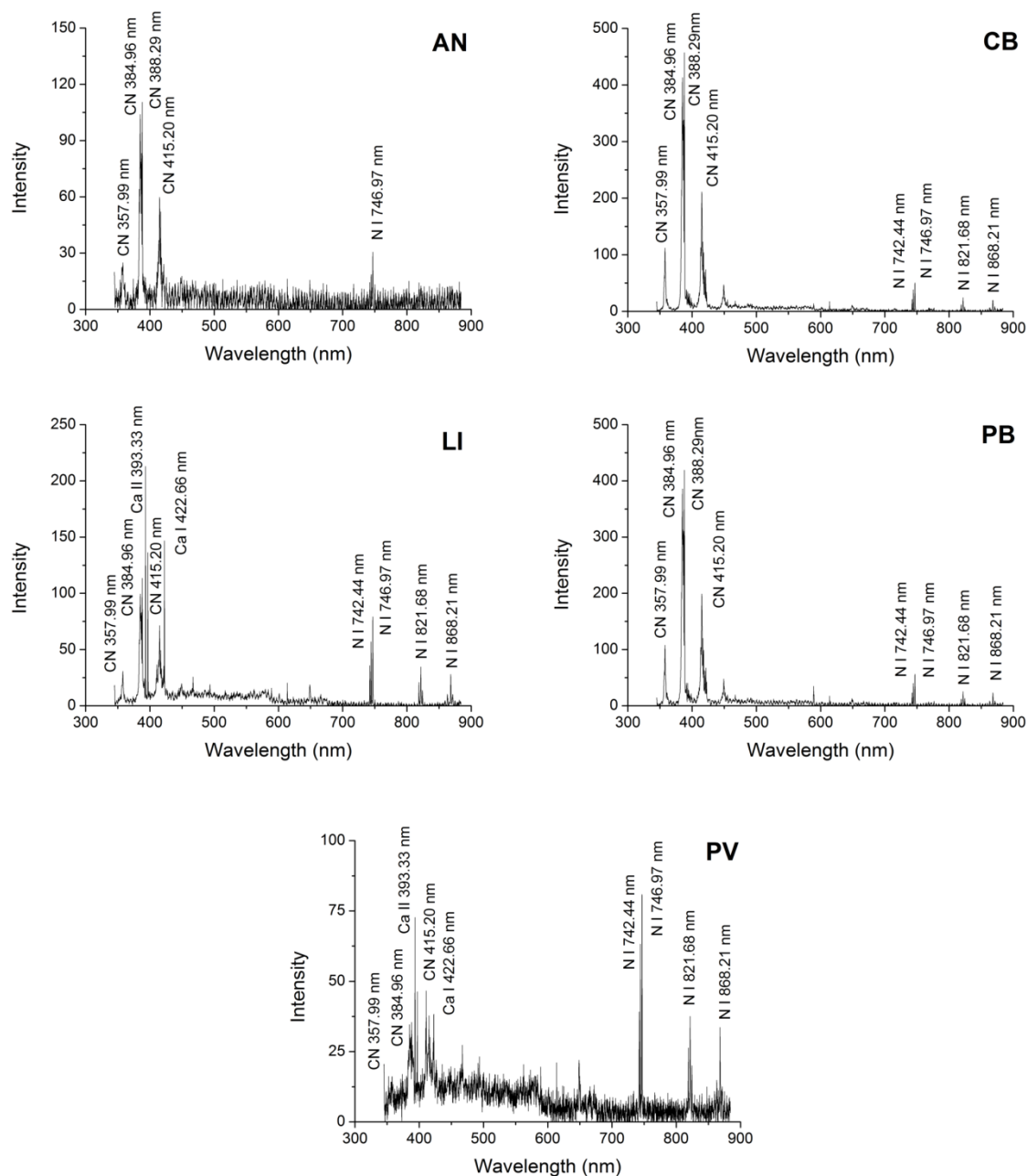


Figure 1. Visible range, typical LIBS spectra of the studied coal aerosols (lignite (LI), anthracite (AN), Pécs-vasas brown coal (PV), Polish brown coal (PB), Czech brown coal (CB)).

As it was expected, during the pre-processing of the LIBS data it became evident that only a part of the delivered 500 laser pulses actually hit particles directly or close enough to produce useful spectral information related to the coal aerosol particles. This fraction (*"particle detection frequency"*) can be expressed as the percents of the number of useful (particle) spectra per the total number of laser shots delivered. The presence of CN bands in the spectra was chosen as the indicator, and only those spectra were retained for further processing, which had integral CN band intensities in the range of 380-389 nm over three times of the standard deviation of the integral intensities in the same range of blank spectra.

The particle detection frequency is expected to be related to the concentration. Based on aerosol LIBS literature, it can be assumed that particle breakdown is efficient within a spherical region with a radius of 250 μm around the focal spot of the laser beam (by adopting the methodology described by Carranza and Hahn [45] and using $E \propto d^3$ proportionality). In addition this, the size (mass) of the broken down particle(s) also has to reach a certain level to allow detection (size or mass detection limit).

3.2. Estimation of the particle detection limit

The knowledge of the concentration and average size of the particles for each aerosol sample makes it possible to estimate the particle size detection limit. This calculation is based on the viable assumption that particle detection frequency, under identical conditions, is proportional to the mass concentration of the aerosol. The total mass of the particles within the “blast radius” (a plasma formation sphere with a radius of 250 μm , or $6.54 \cdot 10^{-2} \text{ mm}^3$ volume, imagined around the focal spot of the laser beam within which all particles are assumed to break down completely) has to exceed the minimum mass needed for detection. As Fig. 2. reveals, a linear correlation fairly holds for the samples studied here (the density of carbon was taken as the bulk density of $2.26 \text{ g}\cdot\text{cm}^{-3}$ in these calculations [46]). Based on the data in Fig. 2., the minimum detectable aerosol mass concentration is ca. $600 \text{ pg}\cdot\text{mm}^{-3}$. This estimate suggests that particles only get detected if their characteristic diameter is at least $2.3 \mu\text{m}$ or if they are smaller but their number concentration is large enough to produce a mass concentration exceeding $600 \text{ pg}\cdot\text{mm}^{-3}$. Since the characteristic diameter of the laser ablation generated aerosol particles in the present experiments was smaller than $1 \mu\text{m}$ for all coal samples (cf. Table 1.), detection was granted by their sufficiently high number concentration.

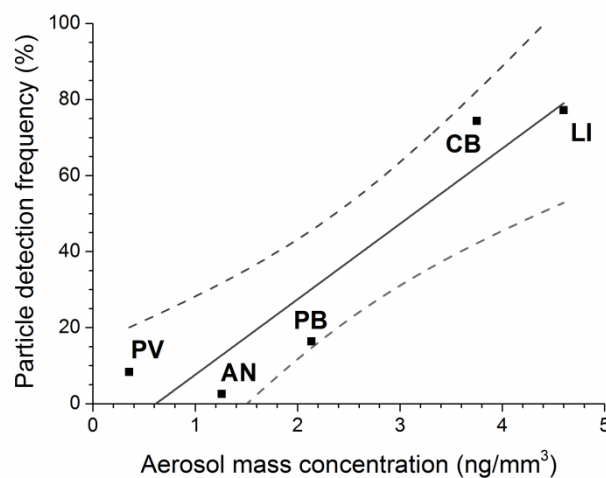


Figure 2. Correlation plot of the particle detection frequency versus the aerosol mass concentration. The dotted curves indicate 90% confidence bands.

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534 It is worth mentioning that the values of the above estimated size and mass
535 concentration detection limits are dependent upon the defined size of the blast radius and a
536 number of experimental factors, including e.g. the laser fluence and coal composition (e.g.
537 volatile matter content, moisture, etc., which influence the breakdown threshold [16]).
538 Nevertheless, the obtained values appear to be reasonable working estimates, as they are
540 comparable to literature values for other particles (e.g. Panne et al. reported about a size
541 detection limit of 0.75 μm for NaCl particles [20] and Carranza et al. about 1 μm for silica
542 particles [45].
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547 **3.3. Discrimination/classification of coal aerosols**

548 ***3.3.1. Sample discrimination by comparative functions***

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551 First, we attempted to use common comparative functions to discriminate between the coal
552 aerosols. We tested three functions, namely linear correlation, sum of squared deviations and
553 overlapping integral, which we had found earlier to be fairly efficient in discrimination
554 applications [16, 31, 34]. The calculations were performed on unscaled data, following spectral
555 masking (data were only retained in the spectral windows of 344-366 nm, 376-430 nm, 460-488
556 nm, 599-652 nm, 730-757 nm, 807-832 nm and 856-884 nm) in order to give more weight to
557 significant spectral features. In spite of these efforts, all the three comparative functions failed to
558 deliver in the present application; there was no significant difference found between the degree
559 of similarity (Q) values and the repeatability was also poor.
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567 ***3.3.2. Sample discrimination by multivariate statistical methods***

568 ***3.3.2.1. Discrimination by multivariate methods combined with MCR-ALS compression***

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570 Data compression is often used in statistical data evaluation in order to decrease the number of
571 variables. It is done mainly for two reasons: *a.)* for some statistical methods, the number of
572 variables has to be small (e.g. LDA/QDA), and *b.)* decreasing the number of variables strongly
573 reduces the size of data matrices and hence the computing time required for the modeling. The
574 approach we chose for data compression was multivariate curve resolution alternating least
575 squares (MCR-ALS), a reasonably modern method, which have already been proved to be an
576 efficient tool (e.g. [34, 38]). Using the Scree-plot approach, it was established that five
577 components (virtual variables) are sufficient to describe most of the variance in the datasets. All
578 calculations were also performed with and without scaling by the autonorm algorithm, which
579 was necessary to compensate for intensity variations in LIBS spectra.
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586 Three modeling approaches were tested: classification tree as well as linear and
587 quadratic discrimination analysis (CLT, LDA, QDA). Training sample datasets consisted of 40,
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randomly selected non-blank spectra of the given sample, according to the procedure described in Section 3.1. The modeling was then performed and repeated 15 times, each time with a different, randomly selected set of 40 spectra. For each statistical model and repetition, the confusion matrix was calculated. The overall accuracy of the model in each repetition was evaluated by calculating the average of the individual accuracies for all five coal aerosol samples (% of correct identifications, in the diagonal of each confusion matrices). Finally, the average and the standard deviation on the accuracy results were calculated (repeatability).

Fig 3. shows the overall discrimination results obtained using the CLT, LDA, QDA methods combined with MCR-ALS data compression. The error bars (standard deviation values) in the graph indicate the repeatability or robustness of the given model; the smaller is the standard deviation, the more reliable or stable the model is.

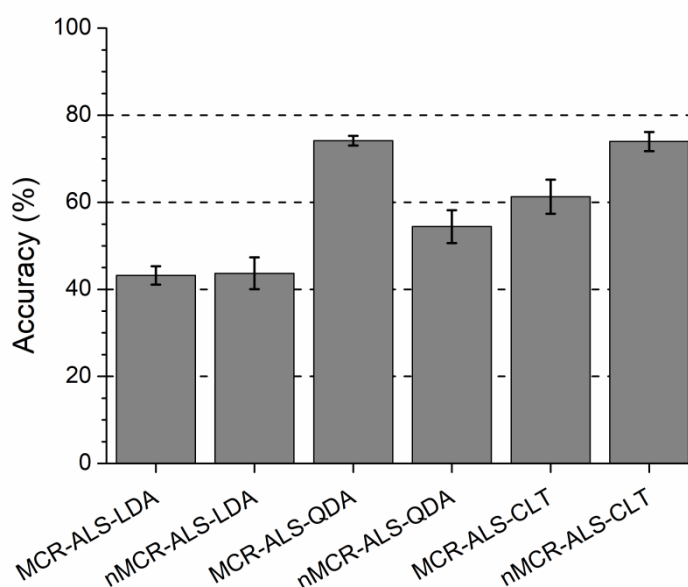


Figure 3. Comparison of the overall (average) accuracy and repeatability (error bars) of the multivariate statistical models used after MCR-ALS data compression. The 'n' in front of the name of the method indicates that the calculation was performed after an automatic intensity scaling by the autonorm algorithm.

The best overall accuracy of the discrimination was clearly provided by the MCR-ALS-CLT method with normalization (74.0%), and by the MCR-ALS-QDA method without normalization (74.2%), although these values are also not excellent. The other two methods performed poorly; their overall accuracy only amounted to 40-60%. The breakaway of the QDA results from the LDA results is understandable, since the QDA works with nonplanar discrimination surfaces that have more flexibility. The CLT method clearly seems to be the most adaptable. The repeatability of the accuracy was good, a few percents only. Normalization was found to produce mixed results: for some methods it gave a weak or strong improvement (MCR-ALS-LDA or MCR-ALS-CLT), but for MCR-ALS-QDA the normalization strongly deteriorated accuracy. Based on these observations, the use of normalization can not be said to be clearly

advantageous. A further investigation of the confusion matrices of the two most accurate compressed methods revealed that although their overall (average) accuracy is very similar, but MCR-ALS-CLT gives a more balanced performance across sample types.

For a numerical illustration of the results, please see Table 2., which presents the confusion matrix for the most accurate model: nMCR-ALS-CLT (the 'n' in front of the name of the method indicates that the data was normalized by the autonorm algorithm). The rows of the matrix correspond to the actual sample classes, whereas the columns correspond to the predicted classes. The values in each cell normally represent the number of samples identified as belonging to the corresponding predicted class based on the statistical evaluation of the measurements. Of course, the ideal case is when all samples are correctly sorted into their corresponding classes, thus the diagonal values of the matrix should be equal to the number of samples (observations) in each actual classes, whereas the off-diagonal values would consequently be zeros (for further information, see e.g. [47]). In Table 2, a color coding and the indication of percentages of identifications were employed to allow a facile overview of performance. As it can be seen, the accuracy (% correct identifications, in the diagonal) is only good (above 80%) for just three samples, and it is fair (around 60%), for two samples. The latter sub-par values are clearly due to frequent false cross classifications between samples CB and PB (in above 20% of the cases) and with AN (above 10%). In other cases, false classifications are rare (below 10%).

In summary, the classification results achievable with MCR-ALS compression were deemed fair, but clearly not adequate for a reliable discrimination.

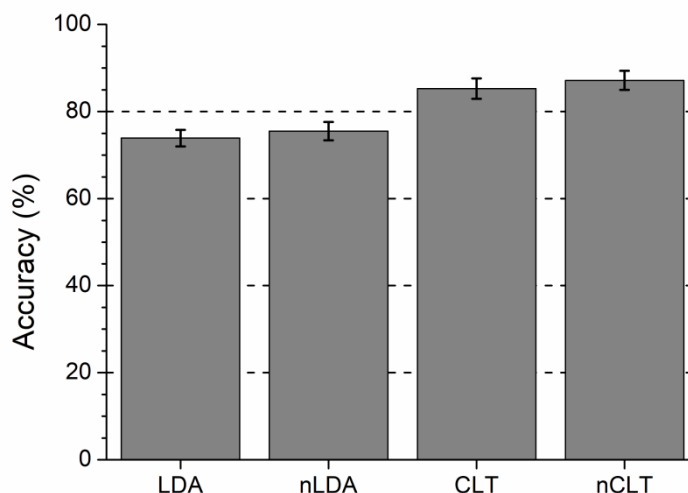
		Predicted class				
		CB	PB	LI	PV	AN
Actual class	CB	57.2 ± 8.2	25.0 ± 9.4	2.3 ± 2.4	3.7 ± 2.7	11.8 ± 5.3
	PB	21.0 ± 7.3	60.3 ± 10.6	1.0 ± 1.8	6.0 ± 3.8	11.7 ± 5.8
	LI	2.2 ± 2.5	1.0 ± 1.6	80.7 ± 8.4	8.8 ± 5.7	7.3 ± 6.0
	PV	0.5 ± 1.9	0.3 ± 0.9	7.7 ± 2.9	81.3 ± 4.7	10.2 ± 5.4
	AN	0.0 ± 3.9	0.0 ± 0.0	2.7 ± 5.1	6.0 ± 7.6	90.3 ± 9.0

Table 2. Confusion matrix for method nMCR-ALS-CLT, containing data for the average accuracy (%) and repeatability from 15 repeated calculations for all samples with randomly selected spectra. Color coding indicates good (green), fair (yellow) and poor (red) performance values in terms of accuracy and repeatability (see the text for more information).

3.3.2.2. Discrimination by multivariate methods without compression

In an attempt to further improve the accuracy of the modelling, we also tested the CLT and LDA methods without data compression. The QDA method could not be tested, as it would require a larger set of spectra (in excess of 2048, that is the number of variables). All calculations were

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711 performed analogously to what was described in the previous section. The overview of the
712 results are shown, with and also without normalization, in Fig 4.
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732 **Figure 4.** Comparison of the overall accuracy and repeatability (error bars) of the multivariate statistical
733 models used without data compression. The 'n' in front of the name of the method indicates that the
734 calculation was performed after an intensity scaling by the autonorm algorithm.
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738 As it can be seen, the accuracy of the modeling greatly improved as a result of not using
739 data compression (cf. Fig. 3). The best overall accuracy was achieved by using the CLT method
740 with normalization; it was 87.2% with a repeatability of 2.2%. Again, it was found that the
741 advantage brought about by using data normalization is only marginal, so its use can be
742 considered as optional. This is somewhat unexpected, because the CLT method employs a set of
743 conditions based on the intensity of spectral lines (Fig 5.). Nevertheless, this corroborates that
744 the repeatability of LIBS spectra of the coal aerosols used here was good. This can be due to two
745 reasons: *a.)* the number concentration was high, thus hundreds of particles were located within
746 the blast radius, and *b.)* the "synthetic" coal aerosol particles were produced by laser ablation
747 under controlled conditions, thus were reasonably similar.
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752 In Table 3., the confusion matrix can be seen for the model CLT that was found to be the
753 most accurate. It is apparent that the performance of this method is good and well balanced: not
754 only that 83.5% to 98.0% accuracy was achieved for 4 out 5 samples, but false classifications are
755 also rare (below 10%), except for sample PB (14.8%). It is also noteworthy that the classification
756 was also successful for aerosol samples that were found to have a mass concentration close to
757 the detection limit estimated in Section 3.2. This can be most probably attributed to that
758 although the detection of the particles for these samples (e.g. PV and AN) was unfrequent and
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 770 weak, but their LIBS spectrum is characteristic enough to facilitate discrimination, e.g. due to
 771 their elemental fingerprint.
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		Predicted class				
		CB	PB	LI	PV	AN
Actual class	CB	83.5 ± 7.5	8.5 ± 6.3	1.0 ± 1.6	1.2 ± 1.3	5.8 ± 3.9
	PB	14.8 ± 9.4	72.7 ± 9.9	1.8 ± 2.0	1.0 ± 1.3	9.7 ± 3.0
	LI	3.8 ± 3.5	0.5 ± 1.4	89.0 ± 4.4	2.7 ± 2.9	4.0 ± 3.1
	PV	0.3 ± 1.3	0.0 ± 0.0	2.3 ± 0.6	92.7 ± 1.1	4.7 ± 0.9
	AN	2.0 ± 3.4	0.0 ± 0.0	0.0 ± 0.0	0.0 ± 0.0	98.0 ± 3.4

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 785 **Table 3.** Confusion matrix for method nCLT, containing data for the average accuracy (%) and
 786 repeatability from 15 repeated calculations for all samples with randomly selected spectra. Color coding
 787 indicates good (green), fair (yellow) and poor (red) performance values in terms of accuracy and
 788 repeatability (see the text for more information).
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790 The use of the classification tree discrimination method is particularly useful in physical
 791 sciences because it is known to carry meaningful information about the composition or
 792 characteristics of the samples [39]. In the present case, the branching conditions are based on
 793 intensities measured at certain locations in the spectra, therefore if the condition indices in the
 794 classification tree are converted to wavelengths (Fig 5.), then information can be revealed about
 795 what spectral lines and hence what elemental components in the coal aerosol particles
 796 contributed to the conditions found by the CLT algorithm. The related spectral lines can be of
 797 small intensity, not easily detectable by the naked eye, but they can give a substantial
 798 contribution to the success of discrimination. In Table 3, we collected a list of elements with a
 799 potential contribution. The list is based on a search in the NIST Atomic Spectroscopy Database.
 800 Only major (C, H, S, O, N) and minor (Na, Mg, K, Ca, P, Ti, Mn, Fe, Al, Si) elements reported as
 801 typically occurring in coal [48], and in the case of the latter group, only reasonably strong
 802 spectral lines were considered. As it can be seen, indeed, there is a potential compliance for
 803 elements like C, H, S, Ca, Na, Si, Fe, Mn, Ti and Al.
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CLT condition (nm)	Potentially contributing elements
358.0	Fe, Ti, Mn
363.8	Fe, Ti
384.4	Ti, Fe, Mn, S
396.8	H, Ca, Fe, Mn
570.0	C, Na, Si, Fe
586.2	Fe, Al
754.6	Fe
768.6	C, Fe, Ti

Table 4. List of elements generally present in coal at major or minor concentration levels that can potentially contribute to spectral intensities leading to branching conditions found by the CLT algorithm.

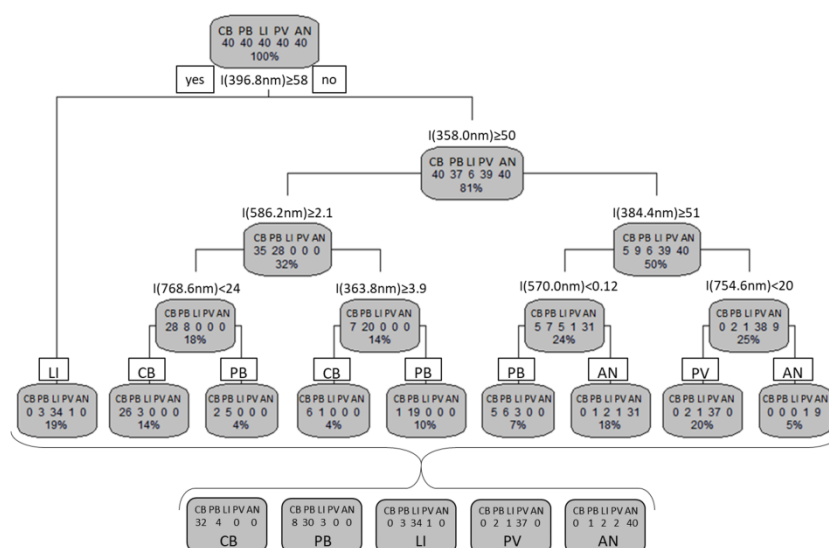


Figure 5. A representative classification tree obtained by the nCLT method. Branching conditions are formulated by the algorithm based on spectral intensities.

4. CONCLUSIONS

In the present work, we demonstrated the capabilities of LIBS for the qualitative analysis of carbonaceous aerosols. By adopting the blast radius methodology, the size detection limit for coal aerosols was also estimated and found to be about 2.3 μm . This means that for detectability, either the diameter of individual coal aerosol particles should be over this diameter or if they are smaller, their number concentration has to be large enough to exceed the equivalent mass concentration (ca. 600 $\text{pg}\cdot\text{mm}^{-3}$). We also demonstrated that the statistical evaluation of LIBS aerosol spectra, especially using multivariate methods, can provide coal type (source) discrimination. The effect of data normalization (by the autonorm algorithm) and data compression (by the MCR-ALS methodology) prior to modeling was also tested and it was found

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888 that while normalization provides marginal improvement, but data compression is generally not
889 beneficial. The best overall performance was showed by the classification tree method without
890 data compression (87.2% accuracy).
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893 Although our work is a laboratory-based feasibility study, but the results demonstrate,
894 for the first time in the literature according to our knowledge, that LIBS is a method that is
895 capable of identifying the source coal type based on the analysis of coal aerosols with good
896 accuracy, therefore it has the potential to become an important on-line tool for the
897 apportionment of carbonaceous aerosols. The estimated 2.3 μm size detection limit indicates
898 that not only PM10 but, with a more sensitive setup also running with a significantly higher laser
899 repetition rate, also the PM2.5 fine fraction of coal aerosols can be detected and characterized by
900 LIBS.
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905 We would also like to mention that achieving an as low as possible mass concentration
906 limit of detection (LOD) value was not our goal. Nevertheless, it can be estimated that with the
907 optimization of the setup (e.g. higher laser fluence, higher solid angle of light collection,
908 spectrometer throughput optimized for the wavelength of the emission line observed, etc.), a
909 100-fold improvement in the mass concentration LOD is possible, especially if the particles are
910 composed of elements easily detectable by LIBS. The feasibility of such an improvement is
911 clearly indicated by several LIBS aerosol studies (e.g. [19, 20, 45]).
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917 Our results and the fact that the involved instrumentation is portable and robust suggest
918 that LIBS may be successfully used in coal combustion flue gas imission studies. The protocol we
919 propose for this application includes, as the first step, that a LIBS spectrum database should be
920 constructed in the laboratory for coal aerosols generated by a laser ablation based system
921 similar to the one we used in our study. Such ablation systems can relatively easily fine-tuned to
922 generate particles either in the sub-micron or micron range. All major coal types available and
923 used as fossil fuels in the area should be incorporated in this database. Second, a discrimination
924 model is built and trained using the spectrum database. Our results indicate that the CLT method
925 is a good candidate statistical approach. Following this, the same LIBS system should be
926 transported to the field, to the location where the imission is planned to be investigated. On
927 the spot, the LIBS spectrometer can be used to perform an on-line, continuous monitoring of a
928 stream of air. In the event of particle (ensemble) detection, an individual classification result
929 would be produced. Reasonably long data collection sequences (ca. 10-30 min) could then
930 provide a near real-time, continuous estimation of the average % coal aerosol composition per
931 source coal type by the statistical evaluation of the classification data produced for each
932 spectrum collected. The collected LIBS spectra can also give information about the elements
933 present in the individually identified particles. This may then be further used as a fingerprint to
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947 identify the emission source (e.g. power or heating facility) if their input fuel elemental
948 composition is known. This protocol can potentially also be extended to incorporate other fossil
949 fuel types as well.
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956
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